

Title Diversity of biological rhythm and food web stability

Author(s) Akihiko Mougi

Journal Biology Letters Volume 17 Issue 2

Published 10 February 2021

URL https://doi.org/10.1098/rsbl.2020.0673

> この論文は出版社版でありません。 引用の際には出版社版をご確認のうえご利用ください。

1	Research Articles
2	
3 4	Diversity of biological rhythm and food web stability
5	
6 7	Akihiko Mougi
8 9	Institute of Agricultural and Life Sciences, Academic Assembly, Shimane University, 1060 Nishikawatsu-cho, Matsue 690-8504, Japan
10	
11	Email: <u>amougi@gmail.com</u>
12	Phone: +81 852 32 6430
13	
14	Number of words: abstract (142), main text (2419)
15	Number of references: 45
16	Number of figures: 2
17	
18	
19 20	

# 21 Abstract

22

23	present a new coexistence theory, i.e., diversity of biological rhythm. Circadian, circalunar, and
24	circannual rhythms, which control short- and long-term activities, are identified as universal phenomena
25	in organisms. Analysis of a theoretical food web with diel, monthly, and annual cycles in foraging
26	activity for each organism shows that diverse biological cycles play key roles in maintaining complex
27	communities. Each biological rhythm does not have a strong stabilizing effect independently but
28	enhances community persistence when combined with other rhythms. Biological rhythms also mitigate
29	inherent destabilization tendencies caused by food web complexity. Temporal weak interactions due to
30	hybridity of multiple activity cycles play a key role toward coexistence. Polyrhythmic changes in
31	biological activities in response to the Earth's rotation may be a key factor in maintaining biological
32	communities.
33	
34	Kowwords: chronobiology biodiversity community persistence, mathematical model
35	Key wor us. enronobology, blourversity, community persistence, mattematical moder
36	
37	
38	
39	
40	
41	
42	
43	
44	
45	
46	
47	
48 40	
50	
51	
52	
53	
54	
55	
56	

How ecosystem biodiversity is maintained remains a persistent question in the field of ecology. Here I

### 57 Introduction

58 Time and space, which are considered fundamental axes in nature, are responsible for the development 59 of ecological communities. In early ecological studies, the spatial dimension was considered the major 60 driver of the dynamics and structure of ecological communities, leading to the development of the field 61 of spatial ecology <sup>1</sup>. However, while it has long been recognized that interactions between species 62 change temporally, research on the temporal dynamics of ecological communities is still at a 63 developmental stage <sup>2,3</sup>. Species interactions can change dynamically over time through environmental 64 fluctuations<sup>4</sup>, adaptive behavior<sup>5</sup>, evolution<sup>6</sup>, and changes in life-history traits<sup>7</sup>, among other 65 mechanisms.

66 In recent years, timing has received increasing attention as climate change threatens to shift 67 phenology, seasonal timing of life-history events<sup>8</sup>, which affects population persistence, and alter 68 community compositions 9-12. Because seasonality causes annual rhythms of species interactions 13, 69 species interactions in active and non-active seasons should differ. For example, predation by a focal 70 species in an active season will be weak or will completely vanish in a non-active season. Although 71 recent studies have explored the roles of phenology in community dynamics and species coexistence 14-72 <sup>20</sup>, they typically consider primarily temporal variations, such as resource fluctuations at a particular 73 trophic level, or focus on a simple community module comprising few species <sup>21–23</sup>. Examining the role 74of phenology on the dynamics and coexistence of diverse species in large ecological communities is a 75 recent challenge <sup>24</sup>.

76 Activity rhythms are not caused by phenology alone. While much research has focused on annual cycles and their role in species coexistence <sup>14–19,22–25</sup>, diel and monthly cycles have received less 77 78 attention. A well-studied empirical system with a diel cycle is the vertical migration of zooplankton <sup>26</sup>, 79 and such diel cycles should work in a wide variety of natural ecosystems because most species exhibit 80 some form of circadian rhythm <sup>27,28</sup>. Nocturnal or diurnal activities are common in many species, and, as 81 with annual cycles, interaction structures can change frequently, even within a day <sup>27,28</sup>. Similarly, a 82 circalunar rhythm leads to monthly activity cycles <sup>29,30</sup>. A notable example is the dramatic change in 83 reproductive activity of marine organisms during a full or new moon, a phenomenon known as a 84 circasyzygic cycle. These remarkable circadian, circalunar, and circannual rhythms, which control 85 organisms' activities over short and long periods, are widespread phenomena and may be considered 86 adaptations to the Earth's rotation and orbit <sup>31</sup>. Nevertheless, how each biological rhythm affects the 87 dynamics and stability of large ecological communities with diverse species remains largely unexplored. 88 Here, I present a polyrhythmic community model incorporating diel, monthly, and/or annual cycles in 89 foraging activity and further describe the role of diversity of biological rhythms in maintaining large 90 ecological communities.

- 91 The proposed model is based on a food web comprising N species, any pair of which is
- 92 connected with the probability C (connectance). Population dynamics are driven by interspecific
- 93 interactions with a non-linear functional response (see Methods). A random food web was assumed to
- 94 reveal the effect of activity rhythms itself in the main text (two other network types are also tested). Diel,
- 95 monthly, and annual cycles in foraging activity are described by sine waves with cycles that have 1-, 30-,
- 96 or 365-day periods. The presence or absence of each rhythm (diel, monthly, and annual) is controlled by
- 97  $\gamma_d$ ,  $\gamma_m$ , and  $\gamma_y$ , respectively ( $\gamma_i = 1$ , where i = day, month, or year, in the presence of a focal rhythm;
- 98 otherwise,  $\gamma_i = 0$ ).
- 99 Controlling the presence or absence of each biological rhythm,  $\gamma_i$ , allows for an examination of 100 the effects of diverse types of biological rhythms in foraging activities on the maintenance of ecological 101 communities. The degree of maintenance was evaluated using community persistence, an index of 102 stability defined as the probability that all species persist for a given time (see Methods for details).

### 103 **Results**

- 104 Consider an extreme case wherein foraging activity does not follow a biological cycle. In a complex 105 community with diverse species, such a system is unlikely to persist according to previous food web 106 models (Fig. 1). Although activity rhythms have been shown to improve the stability of such a fragile 107 system, the stabilization effects that a single rhythm has are weak (Fig. 1). Combining multiple rhythms 108 can increase community persistence. In particular, a system with all available rhythms, daily, monthly, 109 and annual foraging activity cycles, exhibits the highest degree of community persistence (Fig. 1). A 110 variety of biological rhythms can mitigate inherent destabilization due to community complexity (Fig. 111 2). Without biological rhythms, the increase in community complexity (N, C) dramatically decreases 112 community persistence. In contrast, diverse rhythms can prevent the destabilization associated with 113 community complexity.
- The stabilization effect caused by a diversity of biological rhythms can be observed regardless of the network type (Fig. S1). However, a non-linear functional response in foraging is key to exploiting a strong stabilization effect due to the diversity of biological rhythms (Fig. S2, S3). In an extreme case with a linear functional response, the stabilization effect due to a diversity of biological rhythms is found to be weak (Fig. S2).

#### 119 **Discussion**

120 This study describes a novel coexistence mechanism, i.e., diversity of biological rhythm. Diel, monthly,

- 121 and annual cycles in biological activity, known as circadian, circalunar, and circannual rhythms,
- 122 respectively, combine to play a crucial stabilizing role in large ecological communities. Although

123 coexistence is possible under temporal environmental fluctuations, as shown by earlier ecological
 124 theories <sup>21–23</sup>, multiple temporal fluctuations may lead to more stable coexistence.

125 Earlier theoretical studies explored the role of temporal fluctuation in environmental 126 conditions, such as seasonality in species coexistence <sup>22</sup>. For example, in competitive systems, temporal 127 variation in environmental conditions that affect the fitness of competing species can cause temporal 128 niche partitioning and determine coexistence patterns<sup>21</sup>. Timing of species interactions shaped by 129 phenology in the form of match/mismatch among species interaction also plays a key role in driving 130 community dynamics and species coexistence <sup>17, 24</sup>. Periodicity of resource fluctuations can affect competitive coexistence, with longer periodicities promoting species coexistence. Through temporal 131 132 niche partitioning, weakened species interactions can promote species coexistence <sup>23</sup>.

133 The present study shows that activity-level cycles can temporally weaken species interactions 134  $^{32}$ , exerting an inherently stabilizing effect. However, a single activity-level cycle on its own has an 135 insufficient stabilizing effect for the following reasons. Inactive periods in a short cycle are unlikely to 136 be sufficient to ensure recovery in less-abundant and fragile species, and, while a long inactive period 137 may be able to rescue such species, an opposite active period in a long cycle may negate the rescue 138 effect or lead to the extinction of a fragile species. In short, because weak and strong interactions with 139 stabilizing and destabilizing effects are two sides of the same coin in a single active cycle <sup>32</sup>, each active 140 cycle does not necessarily play an effective role in stabilizing community dynamics. However, if 141 multiple activity-level cycles coexist, an entirely different pattern can emerge. Hybridization of multiple 142 activity cycles contributes to a lower distribution of activity level (Fig. S4, S5), i.e., the skew toward low 143 activity reaches a maximum when all cycles are mixed. This skewed distribution of interaction strengths, 144 combined with a rescue effect resulting from a non-linear functional response (lower foraging rates to 145 less-abundant species), plays a key role in stability. This suggests that temporal low activity or "rest" 146 periods during multiple activity cycles can greatly reduce species interaction or cause temporal weak 147 interactions, which is essential for the maintenance of food webs <sup>33</sup>. It is important to note that 148 coexistence is not caused by a storage effect based on competition theory <sup>34</sup> because coexistence occurs 149 even if population dynamics are synchronized (not satisfying species-specific response to environments 150 <sup>34</sup>) (Fig. S6). Additionally, rhythmical changes in activity would be crucial for stability, because random 151 or non-rhythmical changes in activity do not seem to show stabilizing effects (Fig. S7). The temporal 152 weak interactions caused by hybrid activity cycles enable multiple species to coexist by the following 153 mechanism: Even if a short activity-level cycle cannot prevent extinction of a fragile species, a long 154 activity cycle can gradually reverse the species' prospects. This may result from a long-term increase in 155 activity levels or a long-term decrease in activity levels of the species' predators or competitors. Any 156 species that might face extinction because of a long-term unfavorable situation can recover by making

daily and/or monthly changes in its behavior. Activity cycles with different cycle periods may thereforehave complementary effects on stabilizing community dynamics.

159 The complementary effects of multiple activity cycles may have powerful implications for biodiversity conservation. Artificial light at night <sup>35</sup> and temperature increases <sup>36,37</sup> can affect activity 160 161 levels or patterns in daily and monthly cycles, and climate changes, such as global warming, can affect 162 annual active periods <sup>38</sup>. If these changes prolong active phases in biological rhythms at various time 163 scales, activity cycles can be obscured, further preventing stabilization. Because of inherent links between different activity rhythms <sup>39</sup>, alterations in one activity cycle may cause alterations in another. 164 165 The present theory suggests that simultaneous alterations in multiple activity rhythms dramatically 166 decrease the stability of community dynamics. However, empirical tests of this theory are difficult to conduct because creating a natural community without activity cycles is deemed problematic. 167 168 Nevertheless, a possible test of this hypothesis would involve controlling various types of rhythms in 169 environmental conditions, such as light and/or temperature, within a microcosm experiment. Comparing 170 the dynamics of communities before and after removing temporal factors, such as seasonality, using real time-series data analysis <sup>40</sup> offers another option for testing this theory. Furthermore, since the present 171 172 theory emphasizes on the rhythmical change itself, a common stabilizing mechanism may be working in 173 other rhythmical changes, such as a burst-like activity pattern (i.e., short active states between long 174 inactive states) in animal behavior <sup>41</sup>. Future research should be aimed at understanding the relationship 175 between more general diverse rhythm types and higher system-level phenomena. Biological rhythms that 176 evolved in response to the Earth's rotational rhythms by changing the strengths of species interactions in 177 a polyrhythmic manner may play a key role in supporting biodiversity.

#### 178 Methods

179 Consider a random food web in which pairs of species i and j (i, j = 1, ..., N) are connected by a trophic

180 interaction with a probability of C, which is defined as the proportion of realized interaction links L in

181 the possible maximum interaction links  $L_{max}$  of a given network model ( $L = CL_{max}$ ). To examine the

182 generalization of the main result, other types of food webs can be tested (Fig. S1). In a cascade model <sup>42</sup>,

183 for each pair of species i, j = 1, ..., N with i < j, species i never consumes species j, whereas species j

184 may consume species *i*. The maximum link number  $L_{\text{max}}$  is calculated from N(N-1)/2 in both random

and cascade models. In a bipartite model <sup>43</sup>, no interactions occur within the same trophic levels, and

186 species numbers in each of two trophic levels are the same, with  $L_{\text{max}} = (N/2)^2$ . The food web model is

187 defined by an ordinary differential equation:

189 
$$\frac{dX_{i}}{dt} = (r_{i} - s_{i}X_{i} + \sum_{j=1}^{N} A_{ij}X_{j})X_{i}$$
(1)

190

191 where  $X_i$  is the abundance of species *i*,  $r_i$  is the intrinsic rate of change in a species *i*,  $s_i$  is the density-

192 dependent self-regulation of species i, and  $A_{ij}$  is the interaction coefficient between species i and j.

193 Interaction coefficients are defined as  $A_{ij} = e_{ij}\alpha_{ij}$  and  $A_{ji} = -\alpha_{ij}$ , where  $e_{ij}$  (< 1) is the conversion

- 194 efficiency. Next, assume a type II functional response,  $a_{ij} = a_{ij}/(1 + \sum_{k \in \text{resources of species } i} h_{ik} a_{ik} X_k)$ ,
- 195 where  $a_{ij}$  is the consumption rate of species *i* by resource species *j* and  $h_{ij}$  is the handling time.  $a_{ij} =$ 196  $a_{0ij}c_i(t)$ , where  $a_{0ij}$  is the basal consumption rate and  $c_i(t)$  is a time-varying function that represents diel, 197 monthly, and/or annual cycles of foraging activity.
- 198

Each biological cycle is described by a sinusoidal function <sup>44</sup>:

199

200  $c_i(t) = 1 + \gamma_i \sin\{2\pi (t - u_i)/T\},$  (2)

201

202 where  $\gamma_i (= 0 \text{ or } 1)$  is a parameter that determines the presence or absence of each rhythm (see main text) 203 and T and u<sub>i</sub> are the cycle period and timing of activity peak, respectively. T is 1, 30, or 365 in a diel, 204 monthly, or annual cycle, respectively. Note that the time scale of a biological cycle has a day, whereas 205 that of a food web model is assumed to be normalized by a basal resource. In a diel cycle, ui may be 0 or 206 1/2, which is defined as diurnal or nocturnal, respectively. In a monthly cycle,  $u_i$  may be 0 or 30/2, 207 which is defined as a full or new moon, respectively. In an annual cycle,  $u_i$  may be 0 or 365/2, which is 208 defined as summer or winter, respectively. The proportions of species active during bright times (i.e., 209 diurnal, full moon, and summer) are defined as  $p_d$ ,  $p_f$ , and  $p_s$ , respectively; those during night, a new 210 moon, and winter (cold season) are defined as  $1 - p_d$ ,  $1 - p_f$ , and  $1 - p_s$ , respectively. The default setting 211 is  $p_i = 1$  (all species are active during bright times), although relaxation of this assumption does not 212 affect the main result (Fig. S8). The models with either two or all cycles are given by the product of each 213 cycle function with different cycle periods (Eq. 2). For example, in a case with diel and monthly cycles, 214  $c_i(t) = [1 + \sin\{2\pi(t - u_i)\}][1 + \sin\{2\pi(t - u_i)/30\}]$ , and in a case with all cycles,  $c_i(t) = [1 + \sin\{2\pi(t - u_i)/30\}]$ 215  $u_i$ ][1 + sin{2 $\pi$ (t -  $u_i$ )/30}][1 + sin{2 $\pi$ (t -  $u_i$ )/365}]. Note that each function with multiple periods has 216 the same mean value of 1, which allows for the appropriate comparisons with each model, including a 217 null model without a biological cycle. 218 In each iterated simulation, initial species abundance and parameters are randomly selected

219 from the uniform distribution ( $X_i = 0.0$  to 1.0,  $r_i = 0.05$  to 1.0,  $a_{0ij} = 0$  to 0.05,  $e_{ij} = 0.1$  to 0.5). The value

- for  $s_i$  was set to a constant 1.0. A non-zero self-regulation term was used in all species to avoid a
- 221 confounding effect, in which an increase in interspecific links decreases the number of heterotrophic
- 222 species with no potential diet present in the web <sup>45</sup>. From a biological perspective, each species is either
- autotrophic or uses external resources.
- Community persistence <sup>45</sup> was then calculated by measuring the frequencies of all coexistent species ( $X_i > 10^{-3}$  for all *i*) after a sufficiently long time (t = 2 × 10<sup>3</sup>, which corresponded to the time taken for community persistence to reach an asymptote) in 500 runs. Community persistence did not show variance because it is only a probability.

228

#### 229 **References**

- Tilman, D., Kareiva, P. M. & others. Spatial Ecology: The Role of Space in Population Dynamics and Interspecific Interactions. (Princeton University Press, 1997).
- 232 2. Post, E. S. *Time in Ecology*. (Princeton University Press, 2019).
- Wolkovich, E. M., Cook, B. I., McLauchlan, K. K. & Davies, T. J. Temporal ecology in the
  Anthropocene. *Ecol. Lett.* 17, 1365–1379 (2014).
- 4. Stenseth, N. C. *et al.* Ecological effects of climate fluctuations. *Science (80-. ).* 297, 1292–1296
  (2002).
- 5. Fryxell, J. & Lundberg, P. *Individual behavior and community dynamics*. vol. 20 (Springer Science &
  Business Media, 2012).
- 239 6. Thompson, J. N. Rapid evolution as an ecological process. *Trends Ecol. Evol.* **13**, 329–332 (1998).
- 240 7. de Roos, A. M. & Persson, L. *Population and community ecology of ontogenetic development*.
  241 (Princeton University Press, 2013).
- 242 8. Schwartz, M. D. & others. *Phenology: an integrative environmental science*. (Springer, 2003).
- 243 9. Cleland, E. E., Chuine, I., Menzel, A., Mooney, H. A. & Schwartz, M. D. Shifting plant phenology in
  response to global change. *Trends Ecol. Evol.* 22, 357–365 (2007).
- 245 10. Visser, M. E. & Both, C. Shifts in phenology due to global climate change: The need for a yardstick.
  246 *Proc. R. Soc. B Biol. Sci.* 272, 2561–2569 (2005).
- Edwards, M. & Richardson, A. J. Impact of climate change on marine pelagic phenology and trophic
  mismatch. *Nature* 430, 881–884 (2004).
- Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W. & Courchamp, F. Impacts of climate change
  on the future of biodiversity. *Ecol. Lett.* 15, 365–377 (2012).

- 13. Yang, L. H. & Rudolf, V. H. W. Phenology, ontogeny and the effects of climate change on the timing
  of species interactions. *Ecol. Lett.* 13, 1–10 (2010).
- 14. Namba, T. & Takahashi, S. Competitive coexistence in a seasonally fluctuating environment II.
  Multiple stable states and invasion success. *Theor. Popul. Biol.* 44, 374–402 (1993).
- Sakavara, A., Tsirtsis, G., Roelke, D. L., Mancy, R. & Spatharis, S. Lumpy species coexistence arises
  robustly in fluctuating resource environments. *Proc. Natl. Acad. Sci. U. S. A.* 115, 738–743 (2018).
- 16. Nakazawa, T. & Doi, H. A perspective on match/mismatch of phenology in community contexts.
  258 *Oikos* 121, 489–495 (2012).
- Rudolf, V. H. W. The role of seasonal timing and phenological shifts for species coexistence. *Ecol. Lett.* 22, 1324–1338 (2019).
- 26118.Taylor, R. A., Sherratt, J. A. & White, A. Seasonal forcing and multi-year cycles in interacting262populations: lessons from a predator--prey model. J. Math. Biol. 67, 1741–1764 (2013).
- Benincà, E., Ballantine, B., Ellner, S. P. & Huisman, J. Species fluctuations sustained by a cyclic
  succession at the edge of chaos. *Proc. Natl. Acad. Sci. U. S. A.* 112, 6389–6394 (2015).
- 265 20. Godoy, O. & Levine, J. M. Phenology effects on invasion success: Insights from coupling field
  266 experiments to coexistence theory. *Ecology* 95, 726–736 (2014).
- 267 21. Chesson, P. Mechanisms of maintenance of species diversity. *Annu. Rev. Ecol. Syst.* 31, 343–366
  268 (2000).
- 269 22. White, E. & Hastings, A. Seasonality in ecology: Progress and prospects in theory. 2–15 (2018)
  270 doi:10.7287/peerj.preprints.27235v1.
- 271 23. McMeans, B. C., McCann, K. S., Humphries, M., Rooney, N. & Fisk, A. T. Food Web Structure in
  272 Temporally-Forced Ecosystems. *Trends Ecol. Evol.* **30**, 662–672 (2015).
- 273 24. Encinas-Viso, F., Revilla, T. A. & Etienne, R. S. Phenology drives mutualistic network structure and
  274 diversity. *Ecol. Lett.* 15, 198–208 (2012).
- 275 25. Barabás, G., D'Andrea, R. & Stump, S. M. Chesson's coexistence theory. *Ecol. Monogr.* 88, 277–303
  276 (2018).
- 26. Behrenfeld, M. J. *et al.* Global satellite-observed daily vertical migrations of ocean animals. *Nature*576, 257–261 (2019).
- 279 27. Kronfeld-Schor, N. & Dayan, T. Partitioning of Time as an Ecological Resource. *Annu. Rev. Ecol.*280 *Evol. Syst.* 34, 153–181 (2003).
- 281 28. Welch, K. D. & Harwood, J. D. Temporal dynamics of natural enemy-pest interactions in a changing

282		environment. <i>Biol. Control</i> <b>75</b> , 18–27 (2014).
283 284	29.	Raible, F., Takekata, H. & Tessmar-Raible, K. An overview of monthly rhythms and clocks. <i>Front. Neurol.</i> <b>8</b> , 1–14 (2017).
285	30.	Kronfeld-Schor, N. et al. Chronobiology by moonlight. Proc. R. Soc. B Biol. Sci. 280, (2013).
286 287	31.	Dunlap, J. C., Loros, J. J. & DeCoursey, P. J. Chronobiology: biological timekeeping. (Sinauer Associates, 2004).
288 289	32.	Ushio, M. <i>et al.</i> Fluctuating interaction network and time-varying stability of a natural fish community. <i>Nature</i> <b>554</b> , 360–363 (2018).
290 291	33.	Neutel, AM., Heesterbeek, J. A. P. & de Ruiter, P. C. Stability in real food webs: weak links in long loops. <i>Science (80 ).</i> <b>296</b> , 1120–1123 (2002).
292 293	34.	Chesson, P. Multispecies competition in variable environments. <i>Theor. Popul. Biol.</i> <b>45</b> , 227–276 (1994).
294 295	35.	Gaston, K. J., Bennie, J., Davies, T. W. & Hopkins, J. The ecological impacts of nighttime light pollution: A mechanistic appraisal. <i>Biol. Rev.</i> <b>88</b> , 912–927 (2013).
296 297	36.	Körtner, G. & Geiser, F. The temporal organization of daily torpor and hibernation: Circadian and circannual rhythms. <i>Chronobiol. Int.</i> <b>17</b> , 103–128 (2000).
298 299	37.	Lovegrove, B. G. <i>et al.</i> Are tropical small mammals physiologically vulnerable to Arrhenius effects and climate change? <i>Physiol. Biochem. Zool.</i> <b>87</b> , 30–45 (2014).
300 301	38.	Bradshaw, W. E. & Holzapfel, C. M. Genetic response to rapid climate change: It's seasonal timing that matters. <i>Mol. Ecol.</i> <b>17</b> , 157–166 (2008).
302 303	39.	Yerushalmi, S. & Green, R. M. Evidence for the adaptive significance of circadian rhythms. <i>Ecol. Lett.</i> <b>12</b> , 970–981 (2009).
304 305	40.	Deyle, E. R., Maher, M. C., Hernandez, R. D., Basu, S. & Sugihara, G. Global environmental drivers of influenza. <i>Proc. Natl. Acad. Sci. U. S. A.</i> <b>113</b> , 13081–13086 (2016).
306 307	41.	Wearmouth, V. J. <i>et al.</i> Scaling laws of ambush predator 'waiting' behaviour are tuned to a common ecology. <i>Proc. R. Soc. B Biol. Sci.</i> <b>281</b> , 20132997 (2014).
308 309	42.	Chen, X. & Cohen, J. E. Transient dynamics and foodweb complexity in the LotkaVolterra cascade model. <i>Proc. R. Soc. London. Ser. B Biol. Sci.</i> <b>268</b> , 869–877 (2001).
310 311	43.	Thébault, E. & Fontaine, C. Stability of ecological communities and the architecture of mutualistic and trophic networks. <i>Science (80 ).</i> <b>329</b> , 853–856 (2010).

312 313	44.	<ol> <li>Abrams, P. A. When does periodic variation in resource growth allow robust coexistence of competing consumer species? <i>Ecology</i> 85, 372–382 (2004).</li> </ol>					
314 315	45.	Kondoh, M. Foraging adaptation and the relationship between food-web complexity and stability. <i>Science (80 ).</i> <b>299</b> , 1388–1391 (2003).					
316							
317							
318	Supj	plementary material					
319 320	Figur	es S1–S8, Code for Figures					
321	Ethi	cs					
322 323	Not applicable.						
324	Data accessibility						
325 326	No new data was used. The source code for figures has been included in supplementary material.						
327							
328	Competing interests						
329	The author declares no competing interests.						
330							
331	Fun	ding					
332	This work was supported by a Grant-in-Aid for Scientific Research (C) (grant numbers						
333	#20K	06826) from the Japan Society for the Promotion of Science.					
334							
335	Ack	nowledgments					
336	This s	study was supported by a Grant-in-Aid for Scientific Research (C) (#20K06826) from					
337	the Ja	pan Society for the Promotion of Science. The author appreciates the support from the					
338	Facul	ty of Life and Environmental Sciences at Shimane University for providing financial					

339	aid to	publish	this	article.
-----	--------	---------	------	----------

## 341 Author contributions

342	AM	performed	all	research	and	wrote	the	manuscri	pt.

## 348 Figure legends

**Fig. 1**. Effects of foraging activity cycles on stability. Each circle represents a different

350 model with or without an activity cycle. D, M, and Y represent diel, monthly, and annual

351 cycles, respectively. DM, DY, MY, and DMY represent combinations of the cycles. A

352 random model was assumed. N = 50 and C = 0.5.  $h_{ij} = 5$ .

Fig. 2. Relationships between food web complexity and stability. (a) Effects of species richness. C = 0.5. (b) Effects of connectance. N = 50. Red and blue represent cases with or without activity cycles, respectively.  $h_{ij} = 5$ .